

# Switching toxin production on and off: intermittent microcystin synthesis in a *Microcystis* bloom

Susie A. Wood,<sup>1\*</sup> Andreas Rueckert,<sup>2</sup>  
David P. Hamilton,<sup>2</sup> S. Craig Cary<sup>2,3</sup> and  
Daniel R. Dietrich<sup>4</sup>

<sup>1</sup>Cawthron Institute, Nelson, New Zealand.

<sup>2</sup>Department of Biological Sciences, University of Waikato, Hamilton, New Zealand.

<sup>3</sup>College of Marine and Earth Studies, University of Delaware, Newark, DE, USA.

<sup>4</sup>Faculty of Biology, University of Konstanz, Konstanz, Germany.

## Summary

**Toxic cyanobacterial blooms are increasing in prevalence. Microcystins are the most commonly produced cyanotoxin. Despite extensive research the variables regulating microcystin production remain unclear. Using a RT-QPCR assay that allowed the precise measurement of *mcyE* transcriptional gene expression and an ELISA that enabled small changes in total microcystin concentrations to be monitored, we demonstrate for the first time that microcystin production is not always constitutive and that significant up- and downregulation in microcystin synthesis can occur on time scales of 2–6 h. Samples were collected over 3 days from a small eutrophic lake during a dense microcystin-producing *Microcystis* bloom. *McyE* gene transcripts were detected in only four out of 14 samples. Vicissitudes in both microcystin quotas and extracellular microcystin levels corresponded with changes in *mcyE* expression. During the period of exalted microcystin synthesis *Microcystis* sp. cell concentrations increased from 70 000 cells ml<sup>-1</sup> to 4 000 000 cells ml<sup>-1</sup>. These data provide compelling evidence that changes in *Microcystis* cell concentrations influence microcystin production.**

## Introduction

Anthropogenic eutrophication, catchment modification and climate change have been linked to a global intensification of cyanobacterial blooms (Paerl and Huisman, 2008; Conley *et al.*, 2009). The cyanobacteria responsible

for these blooms commonly produce toxic compounds and contact with, or consumption of contaminated water poses a serious health risk to humans, livestock, pets and wildlife (Codd *et al.*, 2005). Toxic blooms have resulted in immeasurable costs to tourism, agriculture, farming and human health worldwide.

Of the known cyanotoxins the hepatotoxic microcystins are the most notorious. Microcystins, of which more than 80 different congeners have been identified (Zurawell *et al.*, 2005), are found ubiquitously worldwide. They are well known for their role in the human fatalities in 1996 when patients of dialysis clinics in Brazil were treated intravenously with microcystin-contaminated water (Azevedo *et al.*, 2002). Microcystins irreversibly inhibit eukaryotic serine/threonine protein phosphatases (e.g. 1 and 2a; MacKintosh *et al.*, 1990; Honkanen *et al.*, 1991) resulting in liver disease as well as nephro- and neurotoxicity (Feurstein *et al.*, 2009). Recently, microcystins have been implicated in increased incidences of human liver cancer in China (Yu, 1995) and carcinogenic responses in rodents (Falconer and Buckley, 1989; Falconer, 1991; Nishiwaki-Matsushima *et al.*, 1992; Ito *et al.*, 1997). Microcystins have therefore been evaluated by the International Agency for Research on Cancer of the World Health Organization whereby microcystin-LR was placed into category 2B, i.e. 'probably carcinogenic for humans' (Grosse *et al.*, 2006).

The complex interactions among physical, chemical and biological variables that lead to the proliferation of cyanobacteria in lentic systems have been the topic of many decades of research. There is still only limited knowledge, however, of the variables that regulate toxin synthesis, e.g. microcystins, at a cellular level. Studies on *Microcystis* spp. (the most common microcystin producing and bloom forming species) have shown correlations between microcystin quotas (total intracellular microcystins per cell), and a multitude of physiochemical variables including: nutrients (Orr and Jones, 1998; Lee *et al.*, 2000; Oh *et al.*, 2000), temperature (Van der Westhuizen and Eloff, 1985) and pH (Van der Westhuizen and Eloff, 1983). However, results of these studies are often contradictory and environmental factors usually only induce changes in microcystin quotas three or fourfold (Sivonen and Jones, 1999).

Most studies of the regulation of microcystins have been undertaken *in vitro* using single strains of

correspondence. E-mail susie.wood@cawthron.org.nz; Tel. (+64) 3548 2319; Fax (+64) 3546 9464.

\*For

microcystin producing cyanobacteria. Such studies can allow single variables to be manipulated and effects on microcystin production to be closely monitored (e.g. Orr and Jones, 1998; Oh *et al.*, 2000; Schatz *et al.*, 2007). Studying cyanobacteria in these 'artificial' environments may, however, alter or remove the variables that regulate the production of microcystins. Changes in cyanobacteria maintained in culture for extended periods are also well known, for example, loss of colonial morphology in *Microcystis* (Zhang *et al.*, 2007) and loss of toxin production (Schatz *et al.*, 2005). In contrast, in-lake studies are challenging because of the plethora of interacting biotic and abiotic variables that may be critical in regulating growth of cyanobacteria and microcystin production.

Recent advances in analytical methods (e.g. ELISA and liquid chromatography-mass spectrometry) have allowed accurate monitoring of microcystins (e.g. Fischer *et al.*, 2001; Dell'Aversano *et al.*, 2004). Microcystins are extremely stable compounds (Tsuji *et al.*, 1995; Harada *et al.*, 1996) and although there is strong evidence that microcystins can be actively transported out of cells (Pearson *et al.*, 2004), it is likely that over time there is some continuous accumulation of these compounds within cells (Young *et al.*, 2005). Thus, measuring minor alterations in intra- or extracellular microcystins may be difficult, or impossible, when relying solely on toxin analysis.

Microcystins are synthesized non-ribosomally by a large (55 kb) peptide synthetase and polyketide synthase enzyme complex. The gene cluster involved in microcystin synthesis (*mcyA–J*) has been identified and sequenced (Nishizawa *et al.*, 2000; Tillett *et al.*, 2000), providing the opportunity to employ molecular techniques to study microcystin production at a genetic level. Moreover, the most recently developed pseudo-viral controlled reverse-transcriptase quantitative PCR (RT-QPCR) assay (Rueckert and Cary, 2009) allows for the accurate assessment of *mcyE* gene expression in toxic *Microcystis* and *Anabaena*. In prokaryotic organisms mRNA is continuously degraded during translation (Voet and Voet, 2004). Using the latter RT-QPCR assay, analysis can be undertaken of temporary transcripts with an average half-life of a few minutes, thus allowing measurement of the transcriptional regulation of microcystin production almost simultaneously with the signal that induces it. The RT-QPCR assay described in Rueckert and Cary (2009) includes an internal armoured RNA standard with the same primer binding sites as the *mcyE* target gene, as well as a unique internal probe binding sequence. This technique provides, for the first time, a method that can be used to quantitatively measure microcystin transcriptional gene expression.

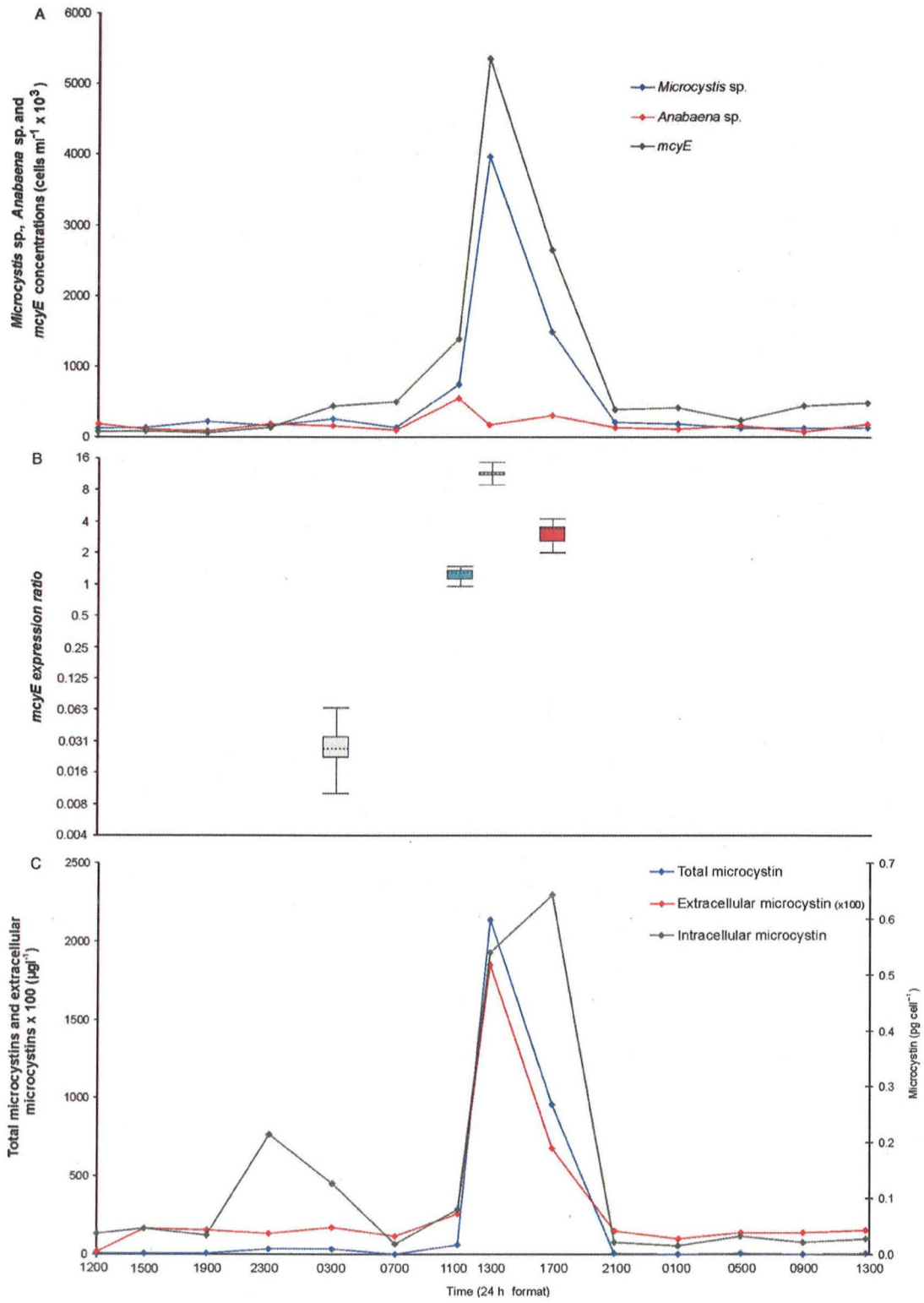
A *Microcystis* bloom comprising of almost entirely toxic genotypes in Lake Rotorua (42°24'05 S, 173°34'57 E,

South Island, New Zealand) provided a valuable opportunity to elucidate how interactions among abiotic and biotic variables influence microcystin synthesis in a cyanobacterial bloom. In this study an RT-QPCR assay (Rueckert and Cary, 2009) and an ADDA-ELISA (Fischer *et al.*, 2001) were used in concert with measurement of a range of physicochemical parameters, microscopic cell counts and QPCR assays to demonstrate for the first time that microcystin synthesis is not always continuous and that changes in *Microcystis* cell concentrations influence microcystin production.

## Results and discussion

Surface water samples were collected from a cyanobacterial bloom in Lake Rotorua (South Island, New Zealand) every 4 h over a 49 h period between 6 and 8 May 2009, hereafter known as days 1, 2 and 3. Using microscopy three cyanobacterial species were identified in surface water samples: *Aphanothece* sp. (<0.5 cell ml<sup>-1</sup>), *Anabaena* sp. *Nova* and *Microcystis* sp. (Fig. 1A). No *Anabaena*-specific *mcyE* genes were detected in the DNA extracted from the preserved filters using QPCR (Vaitomaa *et al.*, 2003), indicating that all microcystins in the samples were produced by *Microcystis* sp. *Microcystis* and non-microcystin producing strains of *Microcystis* can coexist in lake populations (Kurmayer *et al.*, 2002), with this difference attributed to the presence or absence of the microcystin synthetase gene cluster (Nishizawa *et al.*, 2000). QPCR (Vaitomaa *et al.*, 2003) was used to determine the *Microcystis*-specific *mcyE* copy numbers in all samples and the results showed high congruence with the microscopically determined cell concentrations (Fig. 1A), indicating that most, if not all, cells within the population had the ability to produce microcystins.

In this study we used a recently developed RT-QPCR assay (Rueckert and Cary, 2009) to monitor *mcyE* gene transcripts. This assay incorporates a ribonuclease-resistant armoured RNA standard that consists of a non-contagious MS2 bacteriophage virus, in which the genome has been altered to harbour the *mcyE* gene, with a unique internal probe binding sequence for its differential detection. The advantage of this design is that the target and standard sequences are virtually identical, resulting in both sequences undergoing reverse transcription and amplification with the same efficiency. The armoured RNA is used to spike each individual sample, allowing monitoring for the integrity of all nucleic acid processing steps including extraction, reverse transcription, amplification and detection. Using this RT-QPCR assay, *mcyE* gene transcripts were detected in only four out of 14 samples (Fig. 1B). Among the positive samples there were significant up- and downregulations in *mcyE* gene expression over short time periods (2 to 5 h) that



corresponded with changes in microcystin quotas. The highest *mcyE* expression was observed at 1300 h on day 2, and this was 462-fold greater than values from a sample collected 12 h earlier (0300 h) and 12-fold higher than a sample collected just 2 h earlier (Fig. 1B). By 1700 h expression levels had decreased but were still substantially elevated (115-fold) compared with the 0300 sample (Fig. 1B). Previous studies (Kaebernick *et al.*, 2000; Sevilla *et al.*, 2008) have noted variations in the magnitude of *mcy* gene expression; however, these studies have always been undertaken using cultured strains, and changes in *mcy* expression have been small and over longer time frames. Furthermore *mcy* gene expression has always been observed to be continuous rather than extremely variable as observed in our study. Our study is the first to demonstrate that microcystin production is not always continuous over the duration of a cyanobacterial bloom.

Microcystin quotas varied throughout the study period. Analysis using LC-MS (Wood *et al.*, 2010) identified the microcystin variants; -YR, -LR, desmethyl-LR, didesmethyl-LR and WR. When microcystin concentrations were low LC-MS was not suitable for assessing minor changes in microcystin concentrations because individual variants were often below the reliable limits of detection ( $0.2 \mu\text{g l}^{-1}$ ). The ADDA-ELISA (Fischer *et al.*, 2001) measures total ADDA-containing compounds and in this study proved to be an accurate method for assessing small fluctuations in total microcystins (LOD,  $0.02\text{--}0.07 \text{ ng ml}^{-1}$  dependent on variants present). At 1200, 1500 and 1900 h (day 1) the microcystin quotas ranged between  $0.035\text{--}0.046 \text{ pg cell}^{-1}$ . This increased approximately fourfold in the two subsequent samples (2300 h on day 1 and 0300 h on day 2; to  $0.214$  and  $0.126 \text{ pg cell}^{-1}$ , respectively), and then decreased to  $0.019 \text{ pg cell}^{-1}$  at 0700 h before some recovery to  $0.080 \text{ pg cell}^{-1}$  by 1100 h (Fig. 1C). At 1300 and 1700 h (day 2) there was an approximate 15-fold increase in microcystin quotas to  $0.539$  and  $0.644 \text{ pg cell}^{-1}$ , respectively (Fig. 1C), i.e. a 28-fold increase over a 5 h period (1300–1700). To our knowledge no previous studies have shown such a dramatic increase in microcystin quotas of

*Microcystis*, and certainly not within 5 h (Sivonen and Jones, 1999). The observed increase in microcystin quotas could have resulted from an influx of 'more toxic' strains, as some *Microcystis* strains grown in culture are known to always produce greater quantities of microcystins (Bolch *et al.*, 1997). However, the increase in microcystin quotas coupled with the upregulation of *mcyE* provides compelling evidence that this increase is due to some external or internal stimulus. Furthermore, given the continuously high concentrations of *Microcystis* cells present throughout the study ( $> 115\,000 \text{ cells ml}^{-1}$ ) it is likely that any variations in microcystin quotas between strains would be averaged among the population.

The ecophysiological basis for microcystin production has been the topic of many decades of research, with various hypotheses postulated including: defence against grazers (Lüring, 2003), gene regulation (Dittmann *et al.*, 2001), allelopathic interactions (Sukenic *et al.*, 2002), intra-specific regulation (Schatz *et al.*, 2007) and siderophoric scavenging of, and binding to metals such as iron (Martin-Luna *et al.*, 2006; Saito *et al.*, 2008). The data presented in this study provide some further evidence to support and refute some of these hypotheses. The most noteworthy change in a biotic variable that coincided with elevated microcystin quotas and *mcyE* expression was the dramatic increase in the cell concentrations of *Microcystis* between 0700 h and 1300 h (day 2) when there was a 30-fold increase in *Microcystis* concentrations (Fig. 1A). This coincided with a period of gentle onshore winds (data not shown). *Microcystis* colonies are extremely buoyant, commonly forming wind-blown scums (Znachor *et al.*, 2006), and we surmise it was the onshore wind that caused the dramatic increase in cell concentrations. The role of microcystins as a signalling or quorum sensing molecule has recently received increased attention (e.g. Dittmann *et al.*, 2001; Schatz *et al.*, 2007). In our study, as microcystin quotas escalated in Lake Rotorua, there was a corresponding increase in extracellular microcystin concentrations (Fig. 1C), possibly indicating that the microcystins were being exported out of the cell and could be functioning as signalling compounds. However, we have no way of distinguishing between extracellular

**Fig. 1.** Data from surface samples collected at Lake Rotorua (South Island, New Zealand) over a 49 h period (6–8 May 2009). Water samples were preserved (Lugol's Iodine) prior to identification and enumeration and frozen or filtered and frozen, for total and extracellular microcystin analysis. Samples for DNA and RNA extraction were filtered through Whatman GF/C glass microfibre filters and filters were placed immediately in LifeGuard™ Soil Preservation Solution (MO BIO Laboratories, Carlsbad, CA).

A. *Microcystis* sp., *Anabaena* sp. Nova concentrations determined using Utermöhl settling chambers (Utermöhl, 1958). Subsamples (1 ml) were settled and 10 random fields counted at 400× original magnification. *McyE* concentrations were assessed using QPCR (Vaitomaa *et al.*, 2003) and data are an average of two separate analyses.

B. Relative *mcyE* gene expression. Statistical data analysis was performed using the Relative Expression Software Tool (REST 2005 V1.9.12; <http://www.gene-quantification.info>, Pfaffl *et al.*, 2002). The boxes represent the interquartile range or the mid-point of 50% of observations. The dotted line represents the median gene expression. Whiskers represent the minimum and maximum observations.

C. Total and extracellular microcystins, and microcystin quotas as determined by ADDA-ELISA (Fischer *et al.*, 2001). Data represent an average of two separate analyses.

microcystins that have been exported from those that result from cell lysis. Interestingly, although *mcyE* expression and extracellular microcystin levels had decreased by 1700 h on day 2, microcystin quotas remained elevated, perhaps suggesting that cells maintain some microcystins for short periods until their requirement for microcystin abates. Microcystins are extremely stable compounds (Tsuji *et al.*, 1995; Harada *et al.*, 1996) and the rapid decrease in extracellular concentrations (Fig. 1C) is surprising. The most likely explanation is water movement within the lake transported the microcystins out of our sampling area. Bacterial degradation may also have contributed.

The water temperatures measured in Lake Rotorua (8.2–12.5°C; Fig. S1A) are well below those optimal for *Microcystis* growth (Imai *et al.*, 2009). The cooler temperatures may have contributed to the rapid and pronounced changes in *mcyE* expression and microcystin quotas. We speculate that at these low temperatures the cells are already stressed and their response to external or internal stimuli is therefore enhanced.

No other measured abiotic variables changed markedly within the 2 h period (1100–1300 h), when we observed the greatest rate of change in *mcyE* expression and microcystin synthesis. Water temperature increased by only 1.3°C (Fig. S1A) and had changed little by 1700 h when *mcyE* expression had already decreased. Previous culture-based studies have shown only a onefold to twofold increase in microcystin quotas when temperatures have been increased in >5°C increments (e.g. Watanabe and Oishi, 1985). Significant spikes in light intensity were recorded but these did not start until after 1300 h (Fig. S1A), by which time microcystin quotas and *mcyE* expression levels had already increased dramatically. Despite considerable evidence from culture-based studies showing the role of nutrients in regulating microcystin concentrations (e.g. Orr and Jones, 1998; Vezie *et al.*, 2002) changes observed in dissolved nutrients in Lake Rotorua did not correlate with periods of up- or downregulation of microcystins production (Fig. S1B). *Microcystis* and *Anabaena* accounted for almost the entire phytoplankton community; therefore allelopathy does not seem like a viable explanation for the biological role of microcystins in this lake. Although *Anabaena sp. Nova* concentrations did fluctuate (Fig. 1A) these shifts did not align with the marked changes in microcystin synthesis.

Culture-based studies have been the mainstay of research exploring the role and regulation of microcystins. They have typically focused on environmental factors that differ throughout the duration of a cyanobacterial bloom (e.g. light, temperature and nutrients). Often only one parameter is changed while others are maintained at optimal levels. Few, if any, culture-based studies have

changed conditions with the speed and intensity observed in the lake environment, for example, the 30-fold increase in cell concentrations recorded within 6 h in this study. The elucidation of the microcystin biosynthetic pathway has started a new era in the search to understand the function and regulators of microcystins, and this study highlights the benefits of combining highly sensitive molecular and analytical techniques. Using a state-of-the-art technique that allowed the precise measurement of *mcyE* transcriptional gene expression and a sensitive method capable of measuring small changes in total microcystin concentrations, we have demonstrated for the first time that in a cyanobacterial bloom microcystin production is not always constitutive. The significant and rapid up- and downregulation of *mcyE* expression corresponding to variations in microcystin quota, extracellular microcystin levels and cell concentrations adds further evidence to support the hypothesis that microcystins may act as signalling compounds. Increased understanding of the regulation of microcystins in the environment will assist in assessing periods of greatest health risk and may ultimately lead to the development of models enabling real-time predictions of microcystin synthesis to be undertaken.

#### Acknowledgements

This research was funded by the New Zealand Foundation for Research Science and Technology (UOWX0505), a post-doctoral fellowship (CAWX0501) to SAW and a German Science Foundation grant to DRD (DI 698/18). We thank Roel van Ginkel and Michael Boundy (Cawthron Institute) for assistance with LC-MS analysis.

#### References

- Azevedo, S.M., Carmichael, W.W., Jochimsen, E.M., Rinehart, K.L., Lau, S., Shaw, G.R., and Eaglesham, G.K. (2002) Human intoxication by microcystins during renal dialysis treatment in Caruaru-Brazil. *Toxicology* **181**: 441–446.
- Bolch, C.J.S., Blackburn, S.I., Orr, P.T., Jones, G.J., and Grewe, P.M. (1997) Plasmid content and distribution in the toxic cyanobacterial genus *Microcystis* Kützinger ex Lemmermann (Cyanobacteria: Chroococcales). *Phycologia* **36**: 6–11.
- Codd, G.A., Morrison, L.F., and Metcalf, J.S. (2005) Cyanobacterial toxins: Risk management for health protection. *Toxicol Appl Pharmacol* **203**: 264–272.
- Conley, D.J., Paerl, H.W., Howarth, R.W., Boesch, D.F., Seitzinger, S.P., Havens, K.E., *et al.* (2009) Controlling eutrophication: Nitrogen and phosphorus. *Science* **323**: 1014–1015.
- Dell'Aversano, C., Eaglesham, G.K., and Quilliam, M.A. (2004) Analysis of cyanobacterial toxins by hydrophilic interaction liquid chromatography-mass spectrometry. *J Chromatogr A* **1028**: 155–164.

- Dittmann, E., Erhard, M., Kaebernick, M., Scheler, C., Neilan, B.A., Döhren, H., and Börner, T. (2001) Altered expression of two light-dependent genes in a microcystin-lacking mutant of *Microcystis aeruginosa* PCC 7806. *Microbiology* **147**: 3113–3119.
- Falconer, I.R. (1991) Tumor promotion and liver injury caused by oral consumption of cyanobacteria. *Environ Toxicol Water Quality* **6**: 177–184.
- Falconer, I.R., and Buckley, T.H. (1989) Tumour promotion by *Microcystis* sp., a blue-green alga occurring in water supplies. *Med J Aust* **150**: 351.
- Feurstein, D., Holst, K., Fischer, A., and Dietrich, D.R. (2009) Oatp-associated uptake and toxicity of microcystins in primary murine whole brain cells. *Toxicol Appl Pharmacol* **234**: 247–255.
- Fischer, W.J., Garthwaite, I., Miles, C.O., Ross, K.M., Aggen, J.B., Chamberlin, A.R., et al. (2001) Congener-independent immunoassay for microcystins and nodularins. *Environ Sci Technol* **35**: 4849–4856.
- Grosse, Y., Baan, R., Straif, K., Secretan, B., El Ghissassi, F., and Coglianò, V. (2006) Carcinogenicity of nitrate, nitrite, and cyanobacterial peptide toxins. *Lancet Oncol* **7**: 628–629.
- Harada, K.I., Tsuji, K., Watanabe, M.F., and Kondo, F. (1996) Stability of microcystins from cyanobacteria – III. Effect on pH and temperature. *Phycologia* **35**: 83–88.
- Honkanen, R.E., Dukelow, M., Zwiller, J., Moore, R.E., Khatra, B.S., and Boynton, A.L. (1991) Cyanobacterial nodularin is a potent inhibitor of type 1 and type 2A protein phosphatases. *Mol Pharm* **40**: 577–583.
- Imai, H., Chang, K.-H., Kusaba, M., and Nakano, S. (2009) Temperature-dependent dominance of *Microcystis* (Cyanophyceae) species: *M. aeruginosa* and *M. wesenbergii*. *J Plankton Res* **31**: 171–178.
- Ito, E., Kondo, F., Terao, K., and Harada, K. (1997) Neoplastic nodular formation in mouse liver induced by repeated intraperitoneal injections of microcystin-LR. *Toxicol* **35**: 1453–1457.
- Kaebernick, M., Neilan, B.A., Börner, T., and Dittmann, E. (2000) Light and the transcriptional response of the microcystin biosynthesis gene cluster. *Appl Environ Microbiol* **66**: 3387–3392.
- Kurmayer, R., Dittmann, E., Fastner, J., and Chorus, I. (2002) Diversity of microcystin genes within a population of the toxic cyanobacterium *Microcystis* spp. in Lake Wannsee (Berlin, Germany). *Microb Ecol* **43**: 107–118.
- Lee, S.J., Jang, M.H., Kim, H.S., Yoon, B.D., and Oh, H.M. (2000) Variation of microcystin content of *Microcystis aeruginosa* relative to medium N:P ratio and growth stage. *J Appl Microbiol* **89**: 323–329.
- Lüring, M. (2003) Effects of microcystins-free and microcystins containing strains of the cyanobacterium *Microcystis aeruginosa* on growth of the grazer *Daphnia magna*. *Environ Toxicol* **18**: 202–210.
- MacKintosh, C., Beattie, K.A., Klumpp, S., Cohen, P., and Codd, G.A. (1990) Cyanobacterial microcystin-LR is a potent and specific inhibitor of protein phosphatases 1 and 2A from both mammals and higher plants. *FEBS Lett* **264**: 187–192.
- Martin-Luna, B., Sevilla, E., Hernandez, J.A., Bes, M.T., Fillat, M.F., and Peleato, M.L. (2006) Fur from *Microcystis aeruginosa* binds *in vitro* promoter regions of the microcystin biosynthesis gene cluster. *Phytochemistry* **67**: 876–881.
- Nishiwaki-Matsushima, R., Ohta, T., Nishiwaki, S., Suganuma, M., Kohyama, K., Ishikawa, T., et al. (1992) Liver tumor promotion by the cyanobacterial cyclic peptide toxin microcystin-LR. *J Cancer Res Clin Oncol* **118**: 420–424.
- Nishizawa, T., Ueda, A., Asayama, M., Fujii, K., Harada, K.I., Ochi, K., and Shirai, M. (2000) Polyketide synthase gene coupled to the peptide synthetase module involved in the biosynthesis of the cyclic heptapeptide microcystin. *J Biochem* **127**: 779–789.
- Oh, H.M., Lee, S.J., Jang, M.H., and Yoon, B.D. (2000) Microcystin production by *Microcystis aeruginosa* in a phosphorus-limited chemostat. *Appl Environ Microbiol* **66**: 176–179.
- Orr, P.T., and Jones, G.J. (1998) Relationship between microcystin production and cell division rates in nitrogen-limited *Microcystis aeruginosa* cultures. *Limnol Oceanogr* **43**: 1604–1614.
- Paerl, H.W., and Huisman, J. (2008) Blooms like it hot. *Science* **320**: 57–58.
- Pearson, L.A., Hisbergues, M., Börner, T., Dittmann, E., and Neilan, B.A. (2004) Inactivation of an ABC transporter gene, *mcyH*, results in loss of microcystin production in the cyanobacterium *Microcystis aeruginosa* PCC 7806. *Appl Environ Microbiol* **70**: 6370–6378.
- Pfaffl, M.W., Horgan, G.W., and Dempfle, L. (2002) Relative expression software tool (REST<sup>®</sup>) for group-wise comparison and statistical analysis of relative expression results in real-time PCR. *Nucleic Acids Res* **30**: e36.
- Rueckert, A., and Cary, S.C. (2009) Use of an armored RNA standard to measure microcystin synthetase E gene expression in toxic *Microcystis* sp. by reverse-transcription QPCR. *Limnol Oceanogr: Methods* **7**: 509–520.
- Saito, K., Seib, Y., Mikia, S., and Yamaguchi, K. (2008) Detection of microcystin-metal complexes by using cryospray ionization-Fourier transform ion cyclotron resonance mass spectrometry. *Toxicol* **51**: 1496–1498.
- Schatz, D., Keren, Y., Hadas, O., Carmeli, S., Sukenik, A., and Kaplan, A. (2005) Ecological implications of the emergence of non-toxic subcultures from toxic *Microcystis* strains. *Environ Microbiol* **7**: 798–805.
- Schatz, D., Keren, Y., Vardi, A., Sukenik, A., Carmeli, S., Börner, T., et al. (2007) Towards clarification of the biological role of microcystins, a family of cyanobacterial toxins. *Environ Microbiol* **9**: 965–970.
- Sevilla, E., Martin-Luna, B., Vela, L., Bes, M.T., Fillat, M.F., and Peleato, M.L. (2008) Iron availability affects *mcyD* expression and microcystin-LR synthesis in *Microcystis aeruginosa* PCC7806. *Environ Microbiol* **10**: 2476–2483.
- Sivonen, K., and Jones, G. (1999) Cyanobacterial toxins. In *Toxic Cyanobacteria in Water: A Guide to Their Public Health Consequences, Monitoring, and Management*. Chorus, I., and Bartram, J. (eds). London, UK: E&FN Spon, pp. 41–111.
- Sukenik, A., Eshkol, R., Livne, A., Hadas, O., Rom, M., Tchernov, D., et al. (2002) Inhibition of growth and photosynthesis of the dinoflagellate *Peridinium gatunense* by *Microcystis* sp. (cyanobacteria): a novel allelopathic mechanism. *Limnol Oceanogr* **47**: 1656–1663.
- Tillett, D., Dittmann, E., Erhard, M., von Döhren, H., Börner,

- T., and Neilan, B.A. (2000) Structural organization of microcystin biosynthesis in *Microcystis aeruginosa* PCC 7806: an integrated peptide-polyketide synthetase system. *Chem Biol* **7**: 753–764.
- Tsuji, K., Watanuki, T., Kondo, F., Ishikawa, N., Watanabe, M.F., Suzuki, S., *et al.* (1995) Stability of microcystins from cyanobacteria—II. Effect of UV light on decomposition and isomerization. *Toxicon* **33**: 1619–1631.
- Utermöhl, H. (1958) Zur Vervollkommung der quantitativen Phytoplankton Methodik (Towards a perfection of quantitative phytoplankton methodology). *Verh Internat Verein Theor Angew Limnol* **9**: 1–38.
- Vaitomaa, J., Rantala, A., Halinen, K., Rouhiainen, L., Tallberg, P., Mokele, L., and Sivonen, K. (2003) Quantitative real-time PCR for determination of microcystin synthetase E copy numbers for *Microcystis* and *Anabaena* in lakes. *Appl Environ Microbiol* **69**: 7289–7297.
- Van der Westhuizen, A.J., and Eloff, J.N. (1983) Effect of culture age and pH of culture medium on the growth and toxicity of the blue-green alga *Microcystis aeruginosa*. *Z Pflanzenphysiol* **110**: 157–163.
- Van der Westhuizen, A.J., and Eloff, J.N. (1985) Effect of temperature and light on the toxicity and growth of the blue-green alga *Microcystis aeruginosa* (UV-066). *Planta* **163**: 55–59.
- Vezie, C., Rapala, J., Vaitoma, J., Seitsonen, J., and Sivonen, K. (2002) Effect of nitrogen and phosphorus on growth of toxic and nontoxic *Microcystis* strains and on intracellular microcystin concentrations. *Microb Ecol* **43**: 443–454.
- Voet, D., and Voet, J.G. (2004) Prokaryotic mRNA have short lifetimes. In *Biochemistry*, 3rd edn, Vol. 1. Harris, D., and Fitzgerald, P. (eds). New York, NY, USA: Wiley, pp. 98–99.
- Watanabe, M.F., and Oishi, S. (1985) Effects of environmental factors on toxicity of a cyanobacterium (*Microcystis aeruginosa*) under culture conditions. *Appl Environ Microbiol* **49**: 1342–1344.
- Wood, S.A., Heath, M., McGregor, G., Holland, P.T., Munday, R., and Ryan, K. (2010) Identification of a benthic microcystin producing filamentous cyanobacterium (Oscillatoriales) associated with a dog poisoning in New Zealand. *Toxicon* **55**: 897–903.
- Young, F.M., Thomson, C., Metcalf, J.S., Lucocq, J.M., and Codd, G.A. (2005) Immunogold localisation of microcystins in cryosectioned cells of *Microcystis*. *J Struct Biol* **151**: 208–214.
- Yu, S. (1995) Primary prevention of hepatocellular carcinoma. *J Gastroen Hepatol* **10**: 674–682.
- Zhang, M., Kong, F., Tan, X., Yang, Z., Cao, H., and Xing, P. (2007) Biochemical, morphological, and genetic variations in *Microcystis aeruginosa* due to colony disaggregation. *World J Microbiol Biotechnol* **23**: 663–670.
- Znachor, P., Jurczak, T., Komárková, J., Jezberová, J., Maniewicz, J., Klára Kaštovská, K., and Zapomělová, E. (2006) Summer changes in cyanobacterial bloom composition and microcystin concentration in eutrophic czech reservoirs. *Environ Toxicol* **21**: 236–243.
- Zurawell, R.W., Chen, H., Burke, J.M., and Prepas, E.E. (2005) Hepatotoxic cyanobacteria: A review of the biological importance of microcystins in freshwater environments. *J Toxicol Environ Health B* **8**: 1–37.

### Supporting information

Additional Supporting Information may be found in the online version of this article:

**Fig. S1.** Physicochemical data at sampling site at Lake Rotorua (South Island, New Zealand) over a 49 h period (6–8 May 2009).

A. Water temperature and light intensity (approximately 2 cm below water surface) recorded every 5 min (UA-002-08, HOBO®, Onset Computer Corporation, MA).

B. Dissolved reactive phosphorous, ammonium, nitrate and nitrite concentrations determined using standard methods (APHA, AWWA, and WEF, 2005).